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Research paper

Photosynthesis of *Quercus suber* is affected by atmospheric NH_3 generated by multifunctional agrosystems

Marta Pintó-Marijuan^{1,2,6}, Anabela Bernardes Da Silva³, Jaume Flexas⁴, Teresa Dias⁵, Olfa Zarrouk¹, Maria Amélia Martins-Loução⁵, Maria Manuela Chaves¹ and Cristina Cruz⁵

¹Molecular Ecophysiology Lab. (LEM), Instituto de Tecnologia Química e Biológica, Universidade Nova de Lisboa, 2780-901 Oeiras, Portugal; ²Facultat de Biologia, Departament de Biologia Vegetal, Universitat de Barcelona, Av. Diagonal 643, E-08028 Barcelona, Spain; ³Faculdade de Ciências, Centro para a Biodiversidade e Genómica Integrativa e Funcional (BioFIG), Universidade de Lisboa, Campo Grande, 1749-016 Lisboa, Portugal; ⁴Grup de Recerca en Biologia de les Plantes en Condicions Mediterrànies, Departament de Biologia, Universitat de les Illes Balears, Carretera de Valldemossa Km 7.5, E-07122 Palma de Mallorca, Spain; ⁵Faculdade de Ciências, Centro de Biologia Ambiental (CBA), Universidade de Lisboa, Campo Grande, 1749-016 Lisboa, Portugal; ⁶Corresponding author (marta.pinto.marijuan@gmail.com)

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Montados are evergreen oak woodlands dominated by *Quercus* species, which are considered to be key to biodiversity conservation and ecosystem services. This ecosystem is often used for cattle breeding in most regions of the Iberian Peninsula, which causes plants to receive extra nitrogen as ammonia (NH_3) through the atmosphere. The effect of this atmospheric NH_3 ($\text{NH}_{3\text{atm}}$) on ecosystems is still under discussion. This study aimed to evaluate the effects of an $\text{NH}_{3\text{atm}}$ concentration gradient downwind of a cattle barn in a Montado area. Leaves from the selected *Quercus suber* L. trees along the gradient showed a clear influence of the NH_3 on $\delta^{13}\text{C}$, as a consequence of a strong limitation on the photosynthetic machinery by a reduction of both stomatal and mesophyll conductance. A detailed study of the impact of $\text{NH}_{3\text{atm}}$ on the photosynthetic performance of *Q. suber* trees is presented, and new mechanisms by which NH_3 affects photosynthesis at the leaf level are suggested.

Keywords: carbon isotopic discrimination, cork oak, mesophyll conductance, stomatal conductance.

Introduction

As the human population continues to expand, so does the global nitrogen (N) load on the biosphere as a result of increased agricultural activities and fossil fuel emissions, which have nearly doubled since the pre-industrial era (Erisman et al. 2008, Galloway et al. 2008, Rockstrom et al. 2009, Erisman et al. 2010). This relatively recent input of reactive N (all N forms except di-nitrogen, N_2) has a myriad of effects on water, soils, atmosphere and biosphere (Rockstrom et al. 2009). The overall environmental cost of all reactive N loads on the biosphere in Europe has been estimated to be €70–€320 billion per year, outweighing the direct economic benefits to agriculture of the application of reactive N as fertilizer (Sutton et al. 2011). For this reason, besides its scientific importance, the N issue has, at least in Europe, recently gained social, economic and political significance.

On the other hand, managed oak woodlands (Montados, dehesas) dominated by evergreen *Quercus* species, which are often used for cattle breeding in most regions of the Iberian Peninsula, represent a multifunctional integrated agro-silvo-pastoral system. High-quality pastures and agricultural cultivation areas coexist in the same spatial range, providing a good example of a 'human shaped' ecosystem (Bugalho et al. 2011). Consequently, vegetation receives an extra input of N through the atmosphere or the soil (farmyard manure), with likely impacts on plant productivity.

One major implication of the large increase in the reactive N gases delivered to the atmosphere each year is the deposition of this reactive N on the earth's surface. Current total ammonia (NH_3) emissions to the atmosphere are estimated to be 57 Tg N year⁻¹, with the largest contribution coming from

volatilization of animal wastes, 23 Tg N year⁻¹ (Erismán 2012). Both are increasing with time and are expected to be 50% larger by 2050 (Galloway et al. 2004). As atmospheric NH₃ (NH_{3atm}) readily deposits on any surface after emission (Krupa 2003), a large proportion of all emissions are intercepted by plant canopies. Leaf NH₃ uptake may occur through stomata and/or cuticle diffusion (Fernandez and Eichert 2009, Sparks 2009), depending on the relative NH₃ concentrations in the leaf apoplast and the atmosphere. However, actual NH₃ concentrations are dependent on the variables that determine the equilibrium between NH₃ and ammonium (NH₄⁺), mainly water content and pH (Massad et al. 2010).

Several downstream molecular events derived from NH₃ deposition and nutrition have been described as related to C metabolism and energy, e.g., photosynthesis, ion balance, carbohydrate content, respiration rates, the futile cycle of NH₄⁺ influx/efflux (Britto et al. 2001, Ariz et al. 2010, 2011, Balkos et al. 2010, Eichmann et al. 2011), N primary metabolism (Cruz et al. 2006, 2011) and phytohormone and signalling molecules (Barth et al. 2010). When NH₃ enters the cells, it can rapidly be converted into NH₄⁺, affecting all cell metabolism (Cruz et al. 2011). Ammonia/ammonium toxicity was explained either by the uncoupling of photophosphorylation, although not proven for whole organisms (Zhu et al. 2001, Britto and Kronzucker 2002), or as being an inhibitor of photosynthetic oxygen evolution (Sandusky and Yocum 1984). Recently, Polander and Barry (2012) indicated that a water-based hydrogen-bonding network plays a catalytic role in photosynthetic oxygen evolution, and that NH₃ or NH₄⁺ may perturb or disrupt this network, due to the associated increase in bond strength. However, an overall view of the primary steps of NH₄⁺ toxicity in relation to carbon (C) and N metabolism is still lacking, making it difficult to understand plants' physiological responses to NH₃.

We focused on four steps of photosynthetic mechanisms: (i) CO₂ conductance; (ii) leaf structure; (iii) chlorophyll composition and photosystem efficiency; and (iv) enzyme activity. The first step takes into account all the ways that CO₂ reaches chloroplasts and is assimilated by ribulose-1,5-bisphosphate carboxylase oxygenase (Rubisco). Boundary layers, stomata and mesophyll resistance are the main physical interferences to CO₂ assimilation (Flexas et al. 2008). Leaf structure also influences photosynthetic mechanisms: a thicker structure is very often correlated with a higher leaf mass per area (LMA) (Kogami et al. 2001, Poorter et al. 2009) and Hassiotou et al. (2010) positively correlated LMA and mesophyll cell wall thickness ($R^2 = 0.84$; $P_{\text{value}} = 0.01$), which clearly reduces the mesophyll conductance (Flexas et al. 2008, Tosens et al. 2012b). Finally, photosystem efficiency and the highest Rubisco activity affect photosynthesis promoting higher CO₂ assimilation rates (Adams et al. 2013, Parry et al. 2013).

The effect of the NH₃ load on ecosystems is still under discussion. This study examines the effect of a range of NH_{3atm}

concentrations on the photosynthetic performance of a natural forest of *Quercus suber* L., opening up several hypotheses to understand the effects of high concentrations of NH_{3atm} on photosynthesis by woody plants.

Materials and methods

Experimental site

The study site is a cork oak (*Q. suber*) woodland (Montado) in the vicinity of an intensive livestock farm in Portugal (38.744583°N, 8.785405°W). The area has a Mediterranean climate (Rivas-Martínez et al. 2004): dry hot summers, mild rainy winters, average annual atmospheric temperature 17.5 °C and average annual precipitation 600 mm. Meteorological data were collected from the nearest station (38.724293°N, 9.011654°W), at the BA6 base of the Portuguese Air Force. The site does not suffer water limitation due to the presence of groundwater, 1 m below the soil surface. Intensive agricultural and livestock activities take place in the surroundings of the farm, and ~200 beef cows are permanently housed in a single barn measuring 800 m² (resulting in 24 kg N day⁻¹ of extra N input). The soil at the site is sandy alluvial with wind deposits. Up to ~20 m from the barn, animals and machines continually disturb the soil by revolving it, while in the cork oak woodland, soil mobilization is minimal.

Atmospheric ammonia concentrations

Sampling locations (20) were chosen from 2 to 252 m along a transect downwind from the cattle barn. Atmospheric NH₃ concentrations at each location were measured using high-sensitivity adapted low-cost passive high absorption (ALPHA) passive diffusion samplers (Tang et al. 2001), placed 2 m above the ground. Ammonium concentrations on the filter papers collected in the ALPHA samplers (two samplers at each point) were determined by a modification of the Berthelot reaction (Cruz and Martins-Loução 2000; Pinho et al. 2012). Controls and calibrations of the quantification were performed as described in Pinho et al. (2012). Most sampling points were located within 200 m of the barn, where the atmospheric NH₃ gradient was expected to be highest (Sutton et al. 1998). [NH_{3atm}] values are the mean of monthly measurements taken during 1 year, expressed in µg m⁻³.

Leaf sampling

From each *Q. suber*, tree leaves were collected and pooled. Pools were composed of at least 25 current-year, fully expanded, sun-exposed (south orientation). Leaves were detached in the middle of the day, immediately frozen in N₂ liquid and stored at -80 °C until biochemical analyses. Two leaf samplings (LSs) were performed: LS1 and LS2. Trees were selected randomly along the transect from the barn to 250 m away. In LS1, 20 different cork oak trees were selected and in LS2 eight to nine trees.

Leaf NH_4^+ concentration

Each pool of leaves from LS1 was washed with deionised water three times for 30 s, and the water was discarded. Ammonium was determined in extracts prepared by grinding the leaves with liquid N, then adding distilled water at a ratio of 1 g fresh weight:10 ml. Extracts were centrifuged at 5000g at 4 °C and the absorbance of the supernatant was measured as described in the 'Atmospheric ammonia concentrations' section.

Leaf stable C isotope ratios

Five to eight milligrams of powdered plant material from each sample collected in LS1 were separately packed in tin capsules and $^{13}\text{C}/^{12}\text{C}$ isotope ratios were determined by isotope ratio mass spectrometry (IsoPrime Isotope Ratio Mass Spectrometer, Micromass-GV Instruments, Cheadle Hulme, UK) as described in [Werner et al. \(2006\)](#). The leaf C isotope ratio was expressed in delta notation as:

$$\delta^{13}\text{C}_{\text{sample}}(\text{‰}) = \left(\frac{R_{\text{sample}}}{R_{\text{standard}}} - 1 \right) \times 1000$$

where $\delta^{13}\text{C}_{\text{sample}}$ is the isotope ratio in parts per thousand (‰), R_{sample} and R_{standard} are the $^{13}\text{C}/^{12}\text{C}$ molar abundance ratios of the leaf material and the Pee Dee Belemnite standard. Isotope ratios were calibrated against international standards: IAEA CH6 (sucrose) and IAEA CH7 (polyethylene) for C isotope ratios and IAEA N1 (ammonium sulfate) for N isotope ratios ([Rodrigues et al. 2011](#)).

Relative water content determination and leaf morphology

On eight trees selected for LS2, the relative water content (%RWC) of leaves and the LMA were measured. The %RWC was obtained from the formula: $(\text{FW} - \text{DW})/(\text{FSW} - \text{DW}) \times 100$, where FW is the fresh weight, FSW is the fresh saturated weight after rehydrating samples for 24 h in the dark at 4 °C and DW is the dry weight after oven-drying samples at 65 °C until constant weight. The leaf area (LA) was measured with a flat-bed scanner (Epson Perfection V30) and processed using the LA Measurement program (Version 1.3, The University of Sheffield, 2003) image analyzer software. The LMA was calculated as DW/LA . Values given are the means of three independent samples per tree (with four leaves per sample).

Gas exchange and chlorophyll fluorescence measurements

Leaf gas exchange measurements coupled to chlorophyll fluorescence were performed between 10:00 and 16:00 on two attached leaves of each of the nine trees selected for LS2, with a portable infrared gas analyzer (LI-6400 system, LI-COR, Inc., Lincoln, NE, USA) using a leaf chamber fluorometer. In order to

determine the saturation light for C assimilation, the net CO_2 assimilation (A) response curves to photosynthetic photon flux density (PPFD) (A/PPFD) were determined. During the measurements, leaves were maintained at 20 °C, and the supplied CO_2 concentration inside the cuvette (C_a) was $400 \mu\text{mol mol}^{-1} \text{CO}_2$. Light intensity during A/PPFD measurements decreased from 2000 to $0 \mu\text{mol m}^{-2} \text{s}^{-1}$ as follows: 2000, 1500, 1000, 750, 500, 250, 100, 50 and $0 \mu\text{mol m}^{-2} \text{s}^{-1}$. Dark respiration (R_n) was measured in dark-adapted leaves on four trees. Under these conditions, *Q. suber* trees showed light-saturated photosynthesis at $1500 \mu\text{mol m}^{-2} \text{s}^{-1}$. As a consequence, to determine $A/\text{intercellular } \text{CO}_2$ concentration (C_i) response curves to the internal CO_2 concentration, the PPFD was maintained at $1500 \mu\text{mol m}^{-2} \text{s}^{-1}$ (with 10% blue light) and C_i started at $400 \mu\text{mol mol}^{-1}$, decreasing stepwise to 200, $100 \mu\text{mol mol}^{-1}$ and then increasing from 100 stepwise to 150, 200, 250, 350, 500, 750, 1000, 1200, 1500, $1750 \mu\text{mol mol}^{-1}$. The leaf temperature was maintained at 20 °C.

Analyses of the measured A/C_i curves allowed the determination of net CO_2 assimilation at saturating C_i (A_{max}). The maximum apparent carboxylation velocity ($V_{\text{c,max,Ci}}$) for the distinct leaves was determined by fitting a maximum likelihood regression below and above the inflexion of the A/C_i response, using K_c , K_o (Michaelis–Menten constants for CO_2 and O_2 , respectively) as described by [McMurtrie and Wang \(1993\)](#). Values (C_a : $350 \mu\text{mol mol}^{-1}$) of stomatal conductance (g_{s350}), CO_2 assimilation rate (A_{350}) and intrinsic water use efficiency ($\text{WUE}_{\text{L350}} = A_{350}/g_{s350}$) at CO_2 close to atmospheric concentration were also calculated.

Chlorophyll fluorescence was determined concomitantly with each gas exchange measurement. From the fluorescence measurements, the actual quantum efficiency of the photosystem II (PSII)-driven electron transport (Φ_{PSII}) was determined according to [Genty et al. \(1989\)](#), and the rate of linear electron transport (ETR) was calculated as described in [Flexas et al. \(2007\)](#). The maximum photochemical efficiency of PSII (F_v/F_m) was estimated using a portable fluorimeter (Mini-PAM; Walz, Effeltrich, Germany) in three to five dark-adapted (30 min) leaves in the same trees at midday, as described by [Pinto-Marijuan et al. \(2007\)](#).

Estimation of mesophyll conductance

From combined gas exchange and chlorophyll a fluorescence measurements, the mesophyll conductance for CO_2 (g_m) was estimated according to [Harley et al. \(1992\)](#) by assuming that the electron transport rates calculated by gas exchange and by fluorescence match in the absence of internal resistance and photorespiration (PR, [Loreto et al. 1992](#)). Calculated values of g_m were used to convert A/C_i curves to A/C_c curves according to the following equation:

$$C_c = C_i - \left(\frac{A}{g_m} \right)$$

Night respiration (R_n) was used in the experiments as a proxy for R_d by dividing R_n by 2 (Villar et al. 1995, Niinemets et al. 2005). The method of Farquhar et al. (1980) as described in Galle et al. (2011) was used for fitting A/C_c curves, from which g_m at ambient CO_2 concentration ($g_{m,350}$) was estimated and the maximum velocity of carboxylation by Rubisco (V_{c,max_Cc}) was calculated according to Bernacchi et al. (2002).

Rubisco activity and chloroplastic pigments quantification

Rubisco was extracted from *Q. suber* leaves from LS2. Enzyme extraction and determination of total activity (v_t) by the incorporation of $^{14}CO_2$ into acid-stable products at 25 °C in an O_2 -free medium were carried out according to Carmo-Silva et al. (2010). Spectrophotometric measurements to quantify the chloroplastic pigments were performed according to Lichtenthaler (1987).

Statistical analysis

To evaluate the effect of N on the measured traits, the mean values obtained for each parameter on each *Q. suber* tree were compared by a linear regression analysis. R^2 and P_{values} were calculated using the SPSS 15.0 statistical package (SPSS, Inc., Chicago, IL, USA). They are represented in the text and figures as follows: $P \leq 0.01$ (**); $0.01 < P \leq 0.05$ (*); $0.05 < P \leq 0.1$ (*). Statistical graphs were prepared using SigmaPlot 10.0 (Systat Software, Inc., San Jose, CA, USA).

Results

Leaf NH_4^+ concentrations increased according to the distance from the barn

Atmospheric ammonia concentrations were the highest ($>30 \mu g m^{-3}$), close to the cattle barn (Figure 1a), but dropped markedly along the first 50 m of the transect to $<10 \mu g m^{-3}$, stabilizing to background values ($<5 \mu g m^{-3}$) beyond 150 m from the barn. Figure 1b shows a good correlation between $[NH_4]_{leaf}$ and distance from the barn, suggesting that the NH_{3atm} from the barn was able to penetrate leaf tissues.

^{13}C Discrimination along the NH_{3atm} gradient

Leaves from trees near the barn were more enriched in ^{13}C (less negative $\delta^{13}C$ values) than those sampled further from the barn (Figure 1c). The lowest discrimination values were found in the leaves with the highest $[NH_4]_{leaf}$ (Figure 1c, inset; $R^2 = 0.802^{***}$), indicating that NH_{3atm} was able to penetrate the leaf tissues and alter their physiological processes. The parameters NH_{3atm} , $[NH_4]_{leaf}$ and $\delta^{13}C$ were not linearly correlated with the distance from the barn, all decreased exponentially. On the other hand, $\delta^{13}C$ was linearly and positively correlated with $[NH_4]_{leaf}$. When both parameters were compared with photosynthetic characterization, $\delta^{13}C$ always gave the best correlations. This high correlation between $\delta^{13}C$ and NH_{3atm} thus allowed the use of $\delta^{13}C$ as an indicator of the prevailing NH_{3atm} surrounding other sampled trees.

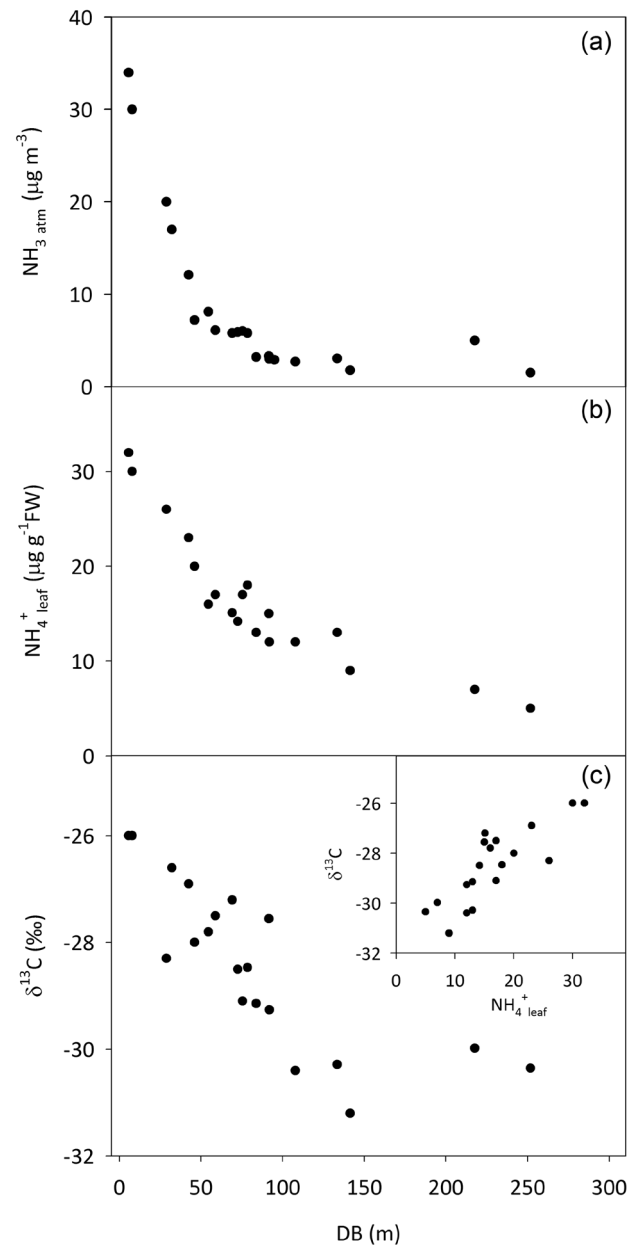


Figure 1. Characterization of the surrounding community along the transect, in relation to the distance from the barn (DB): (a) atmospheric ammonia profile (NH_{3atm} in $\mu g m^{-3}$); (b) ammonium concentration ($[NH_4]_{leaf}$ in $\mu g g^{-1} FW$) in *Q. suber* leaves; and (c) isotopic composition of ^{13}C ($\delta^{13}C$ in ‰) in the same leaves. Inset graph correlates $[NH_4]_{leaf}$ with $\delta^{13}C$ in the sampled leaves.

Two structural parameters were characterized at leaf level (Figure 2): thickness (via LMA) and water content (%RWC). No relation was found between leaf isotopic composition and LMA (Figure 2a; $R^2 = 0.168$). Leaf %RWC was significantly and positively correlated with $\delta^{13}C$ along the transect (Figure 2b; $R^2 = 0.824^{***}$).

Photosynthetic characterization

Assimilation response curves to the internal CO_2 concentration (A/C_i) were determined to identify any effects on physiological

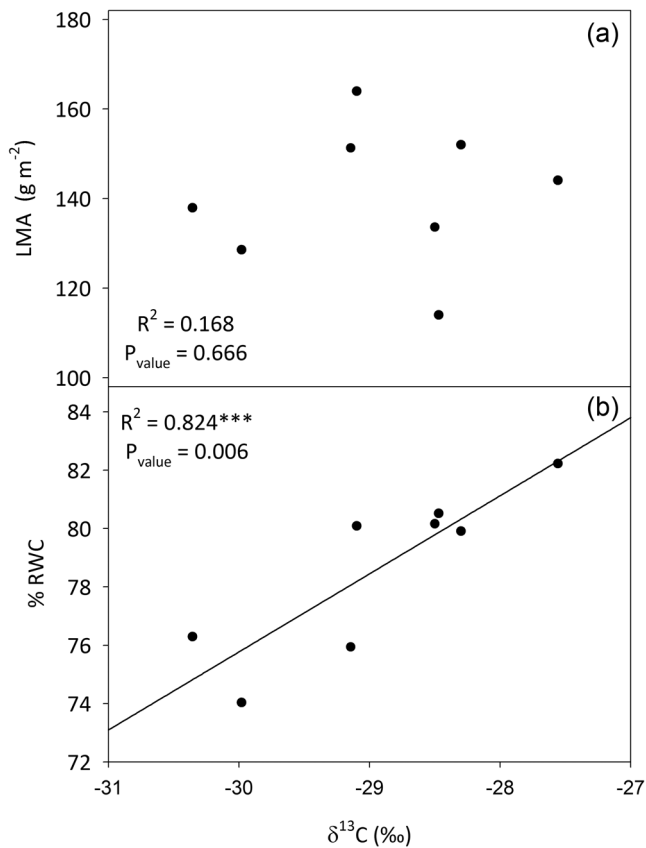


Figure 2. Relationships between leaf isotopic composition ($\delta^{13}\text{C}$, ‰) and physical characteristics of *Q. suber* leaves: (a) leaf thickness expressed as LMA (g DW m^{-2}); and (b) %RWC. The regression coefficients and significance of each relationship are shown in each graph. Asterisks represent the significance as follows: $P_{\text{value}} \leq 0.01$ (***), $0.01 < P_{\text{value}} \leq 0.05$ (**), $0.05 < P_{\text{value}} \leq 0.1$ (*).

mechanisms. Net CO_2 assimilation rate (A_{350} ; Figure 3a: $R^2 = 0.698^{**}$) and stomatal conductance ($g_{s_{350}}$; Figure 3b: $R^2 = 0.654^*$) were negatively correlated with the $\delta^{13}\text{C}$. The strongest negative correlation with ^{13}C discrimination was found with the maximum carboxylation velocity of Rubisco ($V_{c,\text{max}}$; Figure 3c: $R^2 = 0.924^{***}$). This correlation showed a sharp fall in $V_{c,\text{max_Ci}}$ from 140 to $40 \mu\text{mol m}^{-2} \text{s}^{-1}$ along the $\text{NH}_{3\text{atm}}$ gradient. $[\text{NH}_4]_{\text{leaf}}$ affected gas exchange parameters in the studied *Q. suber* trees, reducing the net CO_2 fixation (A_{350}) by reducing both $g_{s_{350}}$ and apparent $V_{c,\text{max_Ci}}$.

Some of the factors expected to affect $V_{c,\text{max_Ci}}$ were as follows: (i) chlorophyll content (Chl *a* + Chl *b*; Figure 4a); (ii) total Rubisco activity expressed per Chl unit (v_t ; Figure 4b); and (iii) photoinhibition, represented by the maximum quantum yield of PSII measured after PSII relaxation in the dark (F_v/F_m ; Figure 4c). However, none of these showed any significant correlation with the C isotopic composition, indicating that they had no effect on $V_{c,\text{max_Ci}}$.

Another factor expected to lower $V_{c,\text{max_Ci}}$ was a decreased mesophyll conductance to CO_2 , which would result in a much lower CO_2 chloroplast concentration ($C_{c_{350}}$) than C_i . A good

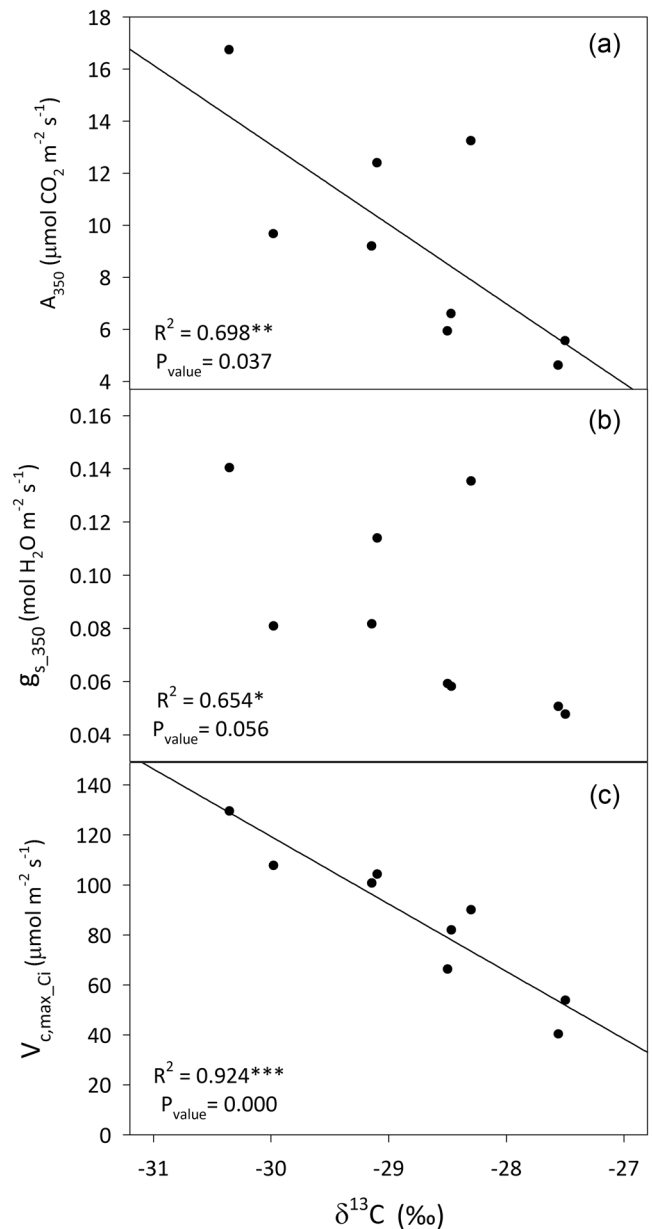


Figure 3. Photosynthetic parameters (calculated from the A/C_i response curves) as a function of isotopic composition ($\delta^{13}\text{C}$, ‰) of *Q. suber* leaves: (a) Net CO_2 assimilation at ambient CO_2 (A_{350} in $\mu\text{mol CO}_2 \text{m}^{-2} \text{s}^{-1}$); (b) stomatal conductance at ambient CO_2 ($g_{s_{350}}$ in $\text{mol H}_2\text{O m}^{-2} \text{s}^{-1}$); and (c) maximum apparent carboxylation velocity ($V_{c,\text{max}}$ in $\mu\text{mol m}^{-2} \text{s}^{-1}$). The regression coefficients and significance of each relationship are shown in each graph. Asterisks represent the significance as follows: $P_{\text{value}} \leq 0.01$ (***), $0.01 < P_{\text{value}} \leq 0.05$ (**), $0.05 < P_{\text{value}} \leq 0.1$ (*).

correlation between $C_{c_{350}}$ and $\delta^{13}\text{C}$ was indeed observed (Figure 5a; $R^2 = 0.555$). The $C_{c_{350}}$ data allowed the recalculation of the maximum velocity of carboxylation by Rubisco based on the chloroplast concentration of CO_2 (Figure 5b; $V_{c,\text{max_Cc}}$). These calculations showed that there were no significant differences between the carboxylation velocities of the studied *Q. suber* trees when the $C_{c_{350}}$ values at the real Rubisco site were taken into account. $C_{c_{350}}$ values also allowed the calculation of the mesophyll

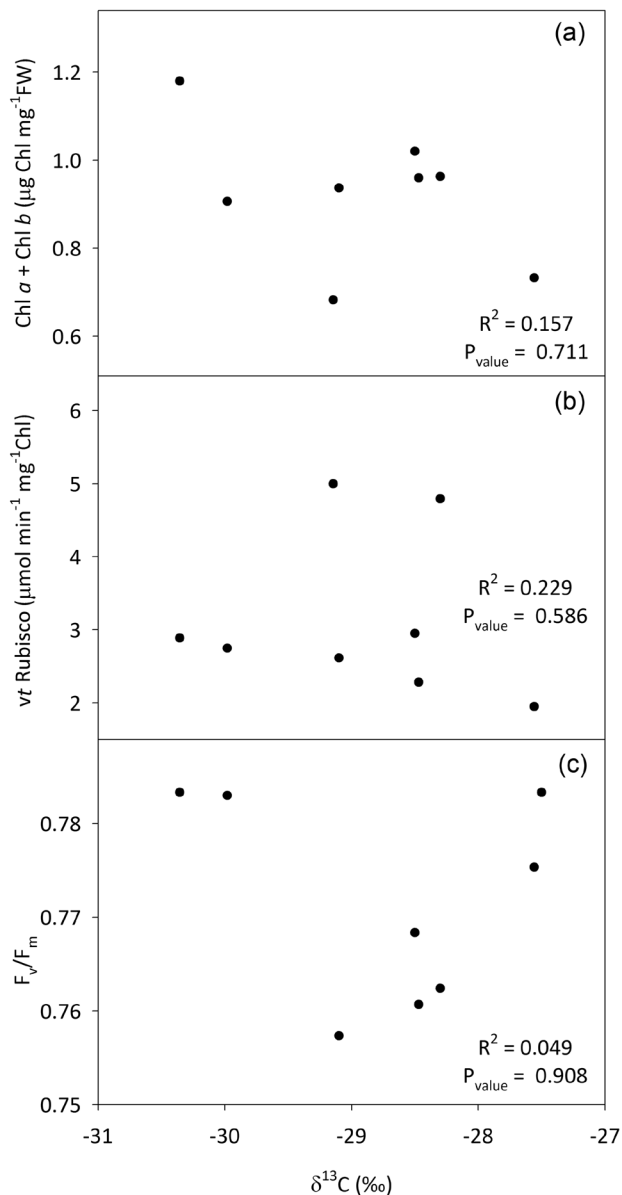


Figure 4. Photosynthetic parameters of *Q. suber* leaves as a function of the leaf isotopic composition ($\delta^{13}\text{C}$, ‰): (a) Total chlorophyll content in μg of pigment mg^{-1} FW; (b) total Rubisco activity (v_t) in μmol of enzyme per min expressed per milligram of total chlorophyll; and (c) maximum photochemical efficiency of PSII (F_v/F_m). The regression coefficients and significance of each relationship are shown in each graph. Asterisks represent the significance as follows: $P_{\text{value}} \leq 0.01$ (***), $0.01 < P_{\text{value}} \leq 0.05$ (**), $0.05 < P_{\text{value}} \leq 0.1$ (*).

conductance at the ambient CO_2 concentration ($g_{m_{350}}$; Figure 5c). $g_{m_{350}}$ was highly correlated with $\delta^{13}\text{C}$ ($R^2 = 0.748^{**}$). The ratios of PR to assimilation calculated at ambient CO_2 concentration (Figure 5d; PR_{350}/A_{350}) showed that the highest PR ($R^2 = 0.647^*$) occurred in trees grown under the highest $[\text{NH}_{3\text{atm}}]$.

Discussion

Montados are traditionally used for free-range cattle breeding, but the application of intensive farming, whereby cattle are

raised in barns at high densities for part of their lives, is spreading. The Montado area studied in the present work is composed of a *Q. suber* forest where an intensive cattle barn was installed, giving rise to a point source of $\text{NH}_{3\text{atm}}$ (Pinho et al. 2011) which generates a downwind decreasing gradient of $[\text{NH}_{3\text{atm}}]$. The movement of NH_3 through the atmosphere is highly dependent on the wind speed, humidity and temperature (Krupa 2003). Since it readily settles on any surface after emission, plant canopies are prone to large deposits of this pollutant (Krupa 2003). In fact, the similarity between observed patterns of $[\text{NH}_{3\text{atm}}]$ and $[\text{NH}_4]_{\text{leaf}}$ along the transect from the barn showed that the cork oak canopies were intercepting $[\text{NH}_{3\text{atm}}]$. After deposition on a leaf's surface, NH_3 may be taken up through stomata and/or cuticle diffusion (Fernandez and Eichert 2009, Sparks 2009), so leaf uptake of gaseous NH_3 by *Q. suber* was expected to be strongly related to stomatal conductance (Gessler et al. 2000, 2002). Our study showed a negative effect of $[\text{NH}_{3\text{atm}}]$ on the stomatal aperture (Figure 3b). The different slopes in Figure 1a (higher slope in $[\text{NH}_{3\text{atm}}]$) and Figure 1b (softer slope of $[\text{NH}_4]_{\text{leaf}}$), due to greater closure of stomata under higher $\text{NH}_{3\text{atm}}$ concentrations, reinforce the hypothesis that $g_{s_{350}}$ affects the entrance of $\text{NH}_{3\text{atm}}$ into the leaf. The diffusion of NH_3 into the leaf apoplast is driven by the concentration gradient between the NH_3 in the air of the sub-stomatal space and the surrounding mesophyll tissue (Krupa 2003). Moreover, the high $[\text{NH}_4]_{\text{leaf}}$ found in our experimental trees ($>30 \mu\text{g g}^{-1}\text{FW}$) revealed some capacity to store NH_4^+ (in the apoplast or in cellular vacuoles), as previously observed in studies conducted at very high NH_3 concentrations (Grundmann et al. 1993, Qiao and Murray 1997).

Leaf isotopic composition data showed that different photosynthetic processes were occurring along the gradient. During photosynthetic CO_2 fixation, fractionation of stable C isotopes occurs and plants consequently become generally depleted in the heavier isotope ^{13}C (Brugnoli and Farquhar 2004). The major steps that discriminate against the heaviest molecule of CO_2 ($^{13}\text{CO}_2$) are biochemical discrimination by the carboxylating enzyme, Rubisco (27–30‰), and the limitations of gaseous diffusive transport from the atmosphere to the carboxylation site (4.4‰) (Farquhar et al. 1982, 1989). Our data showed lower CO_2 discrimination processes at high $\text{NH}_{3\text{atm}}$ concentrations, indicating that it was difficult for CO_2 molecules to reach the chloroplasts on the leaves of the trees that were more exposed to the barn atmosphere. Moreover, our study found CO_2 assimilation (A_{350}) to be strongly correlated with $\delta^{13}\text{C}$ in *Quercus* trees, indicating that the $\text{NH}_{3\text{atm}}$ gradient negatively affected CO_2 fixation. Leaves of trees closer to the barn had lower A_{350} , $g_{s_{350}}$ and very low $V_{c,\text{max-Ci}}$ values (Figure 3); lower stomatal conductance interferes with the movement of CO_2 molecules to the mesophyll, reducing CO_2 availability at Rubisco sites and, therefore, reducing enzymatic discrimination and apparent $V_{c,\text{max}}$. $g_{s_{350}}$ also interferes with the movement of H_2O

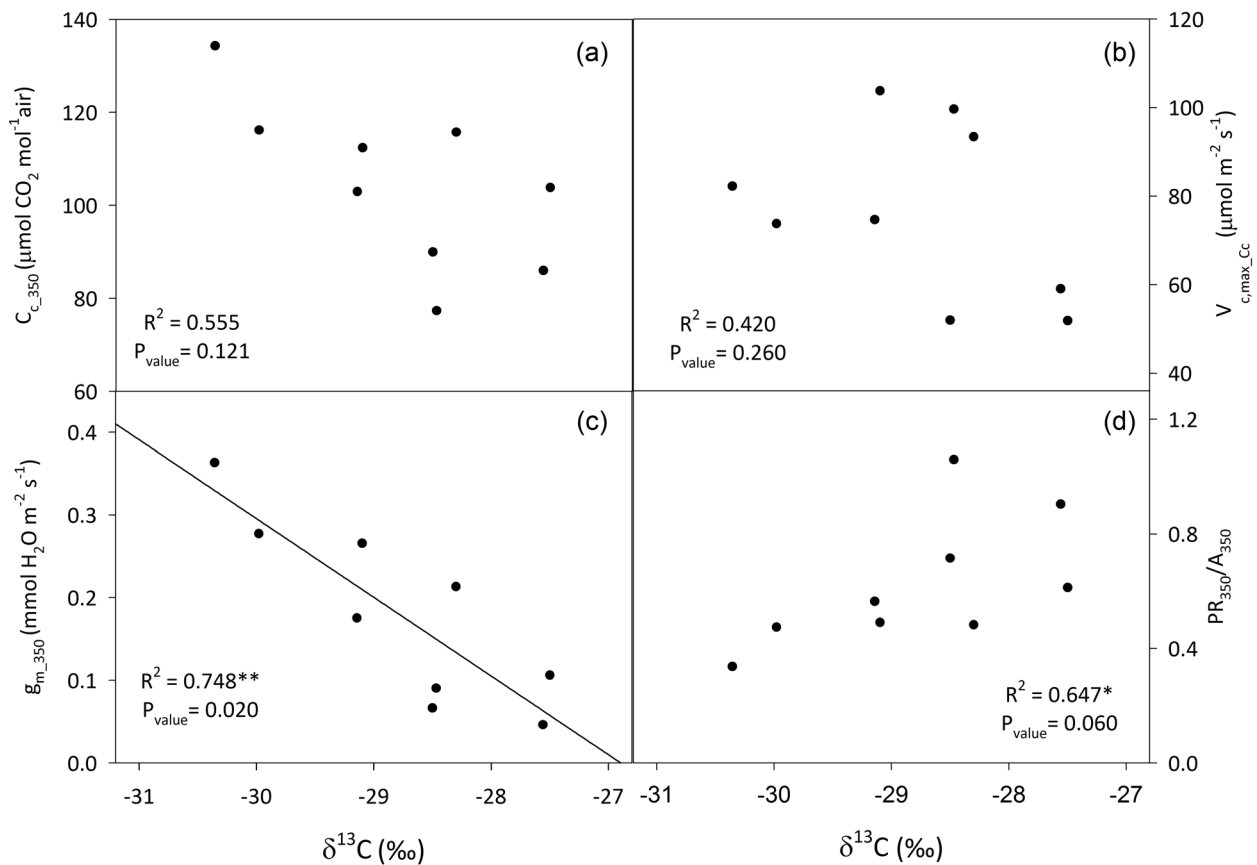


Figure 5. Photosynthetic parameters calculated from A/C_c response curves as a function of isotopic composition ($\delta^{13}\text{C}$, ‰) of *Q. suber* leaves: (a) CO_2 concentration in the chloroplast at ambient CO_2 (C_c in $\mu\text{mol CO}_2 \text{ mol}^{-1} \text{ air}$); (b) maximum apparent carboxylation velocity ($V_{c,\text{max}_C c}$ in $\mu\text{mol m}^{-2} \text{ s}^{-1}$); (c) mesophyll conductance at ambient CO_2 ($g_{m,350}$ in $\text{mmol H}_2\text{O m}^{-2} \text{ s}^{-1}$); and (d) PR divided by net CO_2 assimilation at ambient CO_2 (PR_{350}/A_{350}). The regression coefficients and significance of each relationship are shown in each graph. Asterisks represent the significance as follows: $P_{\text{value}} \leq 0.01$ (***), $0.01 < P_{\text{value}} \leq 0.05$ (**), $0.05 < P_{\text{value}} \leq 0.1$ (*).

molecules from the mesophyll to the atmosphere. Given that water was not a limiting factor on the studied area, we suggest that the strong correlation between $\delta^{13}\text{C}$ and leaf %RWC (Figure 2b) observed among the *Q. suber* trees supported lower stomatal aperture on the trees closer to the barn. In addition, the PR/A ratio at ambient CO_2 concentration (PR_{350}/A_{350} ; Figure 5d) showed that the trees with the highest PR ($R^2 = 0.608^*$) grew under the highest $[\text{NH}_3]$. Our results are in concordance with the findings of Lanigan et al. (2008), who tested the discrimination of glycine decarboxylase against ^{13}C , and concluded that the higher PR implied enriched tissues. Moreover, high PR decreases the apparent conductance of CO_2 diffusion to the chloroplast (Tholen and Zhu 2011). These parameters coherently indicate a photosynthetic limitation on the trees near the barn, a limitation which may be mediated by $[\text{NH}_4]_{\text{leaf}}$.

Several factors that have been previously been indicated to be influenced by $[\text{NH}_3]$ (Ariz et al. 2010, Pompelli et al. 2010) could be the cause of this constraint: lower chlorophyll content, Rubisco concentration or photoinhibition. Chlorophyll content was observed to increase with higher $[\text{NH}_3]$, but data from our experimental site trees (Chl *a* + Chl *b* in Figure 4a) did not

reveal any significant correlation between chlorophylls and C isotopic discrimination, and therefore $\text{NH}_{3\text{atm}}$. The Rubisco content has previously been observed to be higher at high concentrations of NH_4^+ in the growing media (Ariz et al. 2010); however, this study did not find any correlation between variations in v_t (Figure 4b) and $\text{NH}_{3\text{atm}}$. Finally, the maximum quantum yield of PSII photochemistry measured after PSII relaxation in the dark (F_v/F_m ; Figure 4c) might have been affected by $[\text{NH}_3]$ also reducing photosynthetic capacity in high- NH_3 growing plants; however, no such effect was detected along the gradient.

Isotopic discrimination is also related to structural leaf parameters, which can interfere with CO_2 diffusion, such as LMA, an integrative measurement of thickness and density (Fleck et al. 2010). Thickness can limit the ability of CO_2 molecules to reach the chloroplast (Tosens et al. 2012b), contributing to the reduction of discrimination at the Rubisco level. However, our study did not reveal any differences in LMA along the NH_3 gradient, indicating that it was not limited by thicker mesophyll cell walls along the gradient. $V_{c,\text{max}_C c}$ was not a good proxy for Rubisco activity (Figure 3c) when g_m was ignored, as

already demonstrated in several publications (Flexas et al. 2007, 2012). No correlation was found between isotopic composition along the NH_3 gradient and $V_{c,\max_{\text{Cc}}}$ (Figure 5b). These data indicate that Rubisco activity in leaves was stable along the $\text{NH}_{3\text{atm}}$ gradient (Figure 4b). However, a clear and strong correlation ($R^2 = 0.856$; $P = 0.003$) was found between mesophyll conductance for CO_2 (g_m) at ambient CO_2 concentration (Figure 5c; $g_{m,350}$) and $\delta^{13}\text{C}$, indicating that the trees closer to the barn, which had higher concentrations of NH_4^+ in their leaves, presented a high resistance to movement of CO_2 through the mesophyll to the chloroplast lumen.

This is the first time that g_m has been found to be affected by leaf and $\text{NH}_{3\text{atm}}$ concentrations. Here, we present two hypothetical mechanisms of how NH_3 may interfere with $C_{c,350}$ and g_m : (i) anatomical characteristics and (ii) differing aquaporin composition.

First, mesophyll conductance is directly correlated with distinct sclerophyll anatomical characteristics (Tosens et al. 2012b). The strongest sources of g_m variation are the exposed surface area of chloroplasts per unit LA (S_c/S) and the mesophyll cell wall thickness (t_{cw}). Since our study did not observe any changes in LMA, it is not likely that large changes in t_{cw} occurred, which leaves us with the hypotheses that chloroplast structure alters as a function of $[\text{NH}_4]_{\text{leaf}}$. Differences at the chloroplast level have been found in *Moricandia arvensis* grown under an NH_3 source (Winter et al. 1982), and in *Pinus sylvestris* subjected to NH_3 deposition in a fur farm (Back et al. 1997). Structural rearrangements are directly correlated with changes in the g_m data (Tosens et al. 2012a), so we hypothesize that NH_3 could be increasing mesophyll limitations.

Second, aquaporins (AQPs) are water-transporting membrane channel proteins that have also been described as gas channels (Maurel et al. 2008, Herrera and Garvin 2011) ruling CO_2 uptake in both *Nicotiana* and *Arabidopsis* (Uehlein et al. 2008, 2012). Other studies (Kaldenhoff et al. 2008; Flexas et al. 2012) have already described how AQPs facilitate CO_2 mesophyll conductance. It is now known that some AQPs can also transport NH_3 and are widely distributed in cell organelles (Tyerman et al. 2002; Kruse et al. 2006; Hove and Bhavé 2011). Musa-Aziz et al. (2009) described animal/human AQP1 as the only AQP able to transport both CO_2 and NH_3 , while Wudick et al. (2009) suggested that in plants AtTIP1;1 and AtTIP2;1 are located in the chloroplast and are responsible for NH_3 transport, as described by Ferro et al. (2003). The latter authors also mentioned the possible role of AQP in turgor of stomatal guard cells, as previously suggested by MacRobbie (2006). In summary, we hypothesize that a competition at the chloroplast level between CO_2 and NH_3 may occur as a result of: (i) competition inside of the organelle for the same AQP and/or (ii) higher concentrations of NH_3 could induce a greater expression of NH_3 AQP transporters than CO_2 AQP transporters. Whatever the exact mechanisms, a reduced mesophyll

conductance at the highest $[\text{NH}_{3\text{atm}}]$ sites occurs, resulting in a restriction of CO_2 availability at the Rubisco inside the chloroplast, leading to reduced photosynthesis in oak trees.

Conclusions

Atmospheric ammonia is considered to be an air pollutant. Our results indicate that high levels of $[\text{NH}_{3\text{atm}}]$ inhibit the photosynthetic activity of *Q. suber* leaves. This inhibition may be explained by the lower stomatal conductance to CO_2 and the highest resistance of the leaf mesophyll to CO_2 movement, which decreases the availability of CO_2 for carboxylation by Rubisco. This is of particular importance since mesophyll conductance, one of the parameters shown to be affected by $[\text{NH}_{3\text{atm}}]$, is not considered in the majority of models of photosynthetic responses to environmental changes. Although the mechanisms of interaction between NH_3 and CO_2 are still far from understanding, this study opens the prospect of investigating the mechanisms by which NH_3 affects photosynthesis at the leaf level, then scaling this understanding up to models of communities and ecosystems. In consequence, we conclude this study with the prediction that if Montado management continues to allow the inclusion of cattle barns, these should be designed taking into account the effect of high NH_3 concentrations on cork oak productivity. The negative effects of high NH_4^+ concentrations on the receptor ecosystem could be, at least partially, mitigated by creating tree barriers around animal barns.

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Conflict of interest

None declared.

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